

Riparian Meadow Response to Modern Conservation Grazing Management

Kristin M. Oles¹ · Dave A. Weixelman² · David F. Lile³ · Kenneth W. Tate¹ ·
Laura K. Snell⁴ · Leslie M. Roche¹

Received: 25 January 2017 / Accepted: 23 May 2017 / Published online: 2 June 2017
© The Author(s) 2017. This article is an open access publication

Abstract Riparian meadows occupy a small proportion of the public lands in the western United States but they provide numerous ecosystem services, including the production of high-quality forage for livestock grazing. Modern conservation management strategies (e.g., reductions in livestock stocking rates and adoption of new riparian grazing standards) have been implemented to better balance riparian conservation and livestock production objectives on publicly managed lands. We examined potential relationships between long-term changes in plant community, livestock grazing pressure and environmental conditions at two spatial scales in meadows grazed under conservation management strategies. Changes in plant community were not associated with either livestock stocking rate or precipitation at the grazing allotment (i.e., administrative) scale. Alternatively, both grazing pressure and precipitation had significant, albeit modest, associations with changes in plant community at the meadow (i.e., ecological site) scale. These results suggest that reductions in stocking rate have

improved the balance between riparian conservation and livestock production goals. However, associations between elevation, site wetness, precipitation, and changes in plant community suggest that changing climate conditions (e.g., reduced snowpack and changes in timing of snowmelt) could trigger shifts in plant communities, potentially impacting both conservation and agricultural services (e.g., livestock and forage production). Therefore, adaptive, site-specific management strategies are required to meet grazing pressure limits and safeguard ecosystem services within individual meadows, especially under more variable climate conditions.

Keywords Riparian meadows · Public lands · Livestock grazing · Plant community monitoring · Long-term · United States Forest Service · Climate change

Electronic supplementary material The online version of this article (doi:[10.1007/s00267-017-0897-1](https://doi.org/10.1007/s00267-017-0897-1)) contains supplementary material, which is available to authorized users.

✉ Kristin M. Oles
koles@ucdavis.edu

- ¹ Department of Plant Sciences, University of California, One Shields Avenue, Davis, California 95616, USA
- ² USDA Forest Service Pacific Southwest Region, Nevada City, CA 95959, USA
- ³ University of California Cooperative Extension, Susanville, CA 96130, USA
- ⁴ University of California Cooperative Extension, Alturas, CA 96101, USA

1 Introduction

Riparian meadows provide a suite of ecosystem benefits around the world. These diverse ecosystems deliver clean water, flood attenuation, nutrient sequestration, and wildlife habitat (Acreman and Holden 2013; Hatfield and LeBuhn 2007; Norton et al. 2011; Roche et al. 2012). These systems also provide essential forage for grazing livestock to produce protein for a growing human population. Society has strong contemporary expectations for stewardship of grasslands and meadows to balance agricultural goals with social, cultural, and conservation goals in a changing environment (Briske 2011). There is clear evidence that livestock can be managed to conserve and enhance ecosystem services in grazed landscapes (e.g., Davy et al. 2015; Matejkova et al. 2003;

Middleton et al. 2006; Pyke and Marty 2005; Rosenthal et al. 2012). There is also clear scientific evidence that unmanaged, excessive grazing can degrade ecosystems and associated goods and services (Belsky et al. 1999; Eldridge and Greene 1994; Fleischner 1994). In a comprehensive review of the literature, Briske (2011) determined that stocking rate (i.e., livestock grazing pressure) and spatial and temporal distribution of livestock were the primary determinants of agricultural production and conservation outcomes. We argue that to balance multiple goals effectively, public lands grazing strategies must (1) establish and co-value measurable production and conservation objectives; (2) have real-time management action triggers to safeguard ecosystem services; and (3) be adaptive to accommodate spatially and temporally variable, site-specific conditions.

In the western U.S. (the West), millions of hectares of perennial grasslands, shrublands, and forestlands are held in the public domain and managed by state and federal agencies for multiple land uses, including livestock grazing. Riparian meadows account for less than 5% of the area within these landscapes, but provide disproportionate and unique ecosystem services (Bales et al. 2011; Hammersmark et al. 2010; Kuhn et al. 2011). Livestock grazing began on these landscapes during the last half of the 19th century, with numbers steadily rising to an unsustainable peak during World War I. The unregulated grazing pressure of this era led to substantial environmental degradation and impairment of ecosystem services (Sampson and Weyl 1918). In response, in the mid-1920s to early 1950s, livestock numbers were steadily reduced, grazing management areas were established and allotted to specific grazers (known as grazing allotments), and managed grazing systems were implemented to “rehabilitate” upland forage production capacities (Hormay and Evanko 1958). While these changes improved upland health and productivity (Ratliff et al. 1972), grazing plans during the rangeland rehabilitation era (mid-1950s through 1980s) did not co-value riparian areas; in fact, riparian meadows were considered “sacrifice areas” that would necessarily receive heavy livestock grazing to optimize forage harvest across grazing allotments (Meehan and Platts 1978). Thus, the interdependent plant-soil-hydrologic function of riparian meadows continued to degrade during this rangeland rehabilitation era of public lands grazing (Odion et al. 1988). At the root of this degradation was excessive defoliation of riparian meadow vegetation that reduced plant vigor, reproductive capacity, and competitiveness—shifting riparian meadow plant communities from wetland to upland or non-native species. Combined with direct impacts from livestock, these functional plant community changes reduced rooting mass and soil stability, increased soil erosion, impaired hydrologic function, and resulted in riparian desiccation; thus, diminishing a suite of regulating and

supporting ecosystem services (Dwire et al. 2004; Kauffman and Krueger 1984; Kleinfelder et al. 1992; Manning et al. 1989; Micheli and Kirchner 2002; Norton et al. 2011).

During the 1990s, there was a fundamental policy and on-the-ground management paradigm shift in which riparian meadows emerged as critical conservation areas—where production and conservation objectives were co-valued—rather than being relegated to “sacrifice areas” (e.g., Armour et al. 1991; U.S. Government Accounting Office 1988). This shift initiated the modern grazing era that integrated riparian meadow conservation policies and management on US federal public lands (~2000 to present). To meet conservation objectives in this modern era, livestock grazing pressure on federal grazing lands has been reduced by 15% on average across the 11 western states, with extremes of 67% and 36% in Wyoming and California, respectively (Fig. 1). A majority of these reductions were in response to implementation of annual riparian grazing utilization standards. These grazing standards are real-time management action triggers implemented annually to safeguard riparian meadows from excessive livestock utilization and to meet meadow conservation objectives (Freitas et al. 2014). Site-specific adaptive implementation of a suite of practices (e.g., herding, fencing, livestock drinking water developments) can be used to respond to action triggers, meet riparian grazing standards, and balance livestock production and multiple conservation outcomes (George et al. 2011). Modern meadow conservation objectives include maintaining and enhancing diverse, native plant communities resilient to invasion, as well as maintaining and enhancing native perennials and wetland obligate species essential to hydrologic stability.

In the late 1990s, concurrent with the initiation of the modern grazing era, the US Forest Service (USFS) in California initiated a state-wide, long-term meadow vegetation monitoring program (1) to document baseline meadow vegetation prior to the widespread implementation of these new riparian grazing standards; and (2) to examine trends in meadow vegetation following implementation of these riparian grazing standards. Our objective was to utilize data from this state-wide monitoring effort to examine whether long-term vegetation changes were significantly related to livestock grazing pressure and environmental conditions at two scales: meadow scale (i.e., ecological sites measured in hectares) and allotment scale (i.e., administration units measured in square kilometers).

2 Methods

2.1 Study Area

This cross-sectional, longitudinal study was conducted across 138 grazing allotments on 16 USFS managed

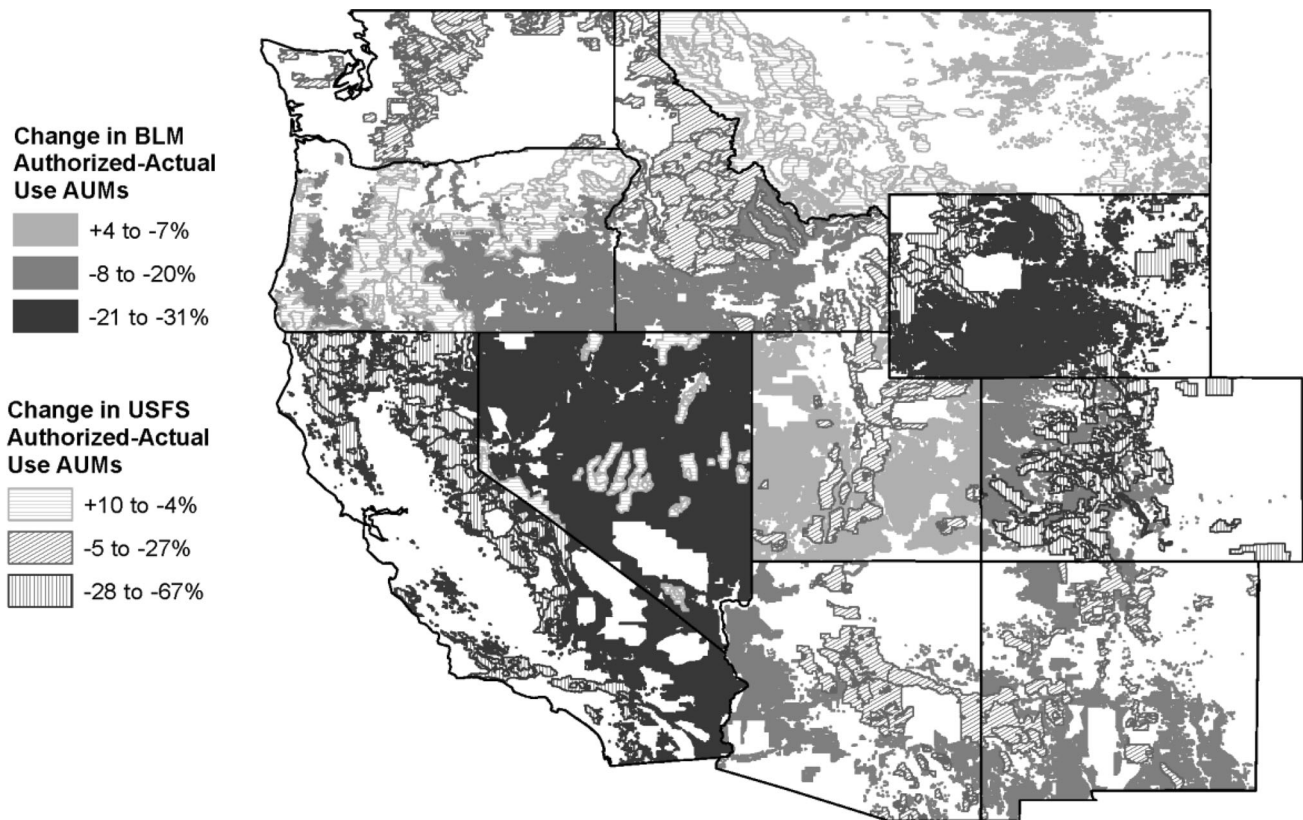


Fig. 1 Change in animal unit months (AUM) on public lands in eleven states in the western U.S. between 2000 and 2015. The *lightest color* represents slight positive to slight negative changes in AUMs. *Darker colors* represent increasingly negative changes in AUMs. *Solid polygons* represent lands administered by the Bureau of Land Management

(BLM). *Hatched polygons* represent lands administered by the U.S. Forest Service (USFS). Data were sourced from BLM and USFS annual reports (Bureau of Land Management 2016; U.S Forest Service 2016)

national forests (Fig. 2) to examine relationships between meadow scale plant community changes and allotment scale environmental conditions and livestock grazing pressure. Each allotment contains at least one long-term (i.e., more than 8 years) meadow plant community monitoring site, with a total 279 sites across the 138 allotments. All plant community monitoring sites were located within riparian meadows fed by a mix of defined stream channels, diffuse overland flow, and emergence of shallow groundwater. The 138 allotments range in elevation from 972 to 3257 m and in area from 12 to 311 km² (Table 1). The majority of precipitation occurs in winter as snow.

From the 138 allotments, a subset of 34 representative grazed allotments (six national forests) were selected to examine relationships between meadow scale plant community changes and meadow scale environmental conditions and livestock grazing pressure (subset of 52 sites – Fig. 2). The subset sites were chosen as a stratified sample from the allotment scale sites to ensure representation of highly variable meadow types. This subset of 52 sites ranges in elevation from 1372 to 2490 m (Table 2). All 138 allotments are located in mountainous watersheds

composed of xeric to mesic forests, with meadows and riparian corridors covering <5% of the landscape (US Forest Service 2016). Meadow plant communities are dominated by perennial grasses and grass-like species. Common species include *Carex nebrascensis* Dewey, *Juncus arcticus* Willd. subsp. *littoralis* (Engelm.) Hultén, and *Symphyotrichum spathulatum* (Lindl.) G.L. Nesom var. *spathulatum*.

2.2 Grazing Management and Riparian Grazing Standards

During the study period (1997–2015), 100 of the 138 study allotments were grazed by beef cattle, 22 by sheep, and 16 allotments received little to no livestock grazing pressure (average annual allotment stocking rates between 0 and 5 animal unit months/km²). An animal unit month (AUM) is the dry weight mass of forage required to sustain one 450 kg cow for a 30-day period (360 kg). Grazing seasons and stocking rates in USFS managed grazing allotments vary depend upon elevation and local climate. Most grazing occurred from May/June through August/September, but

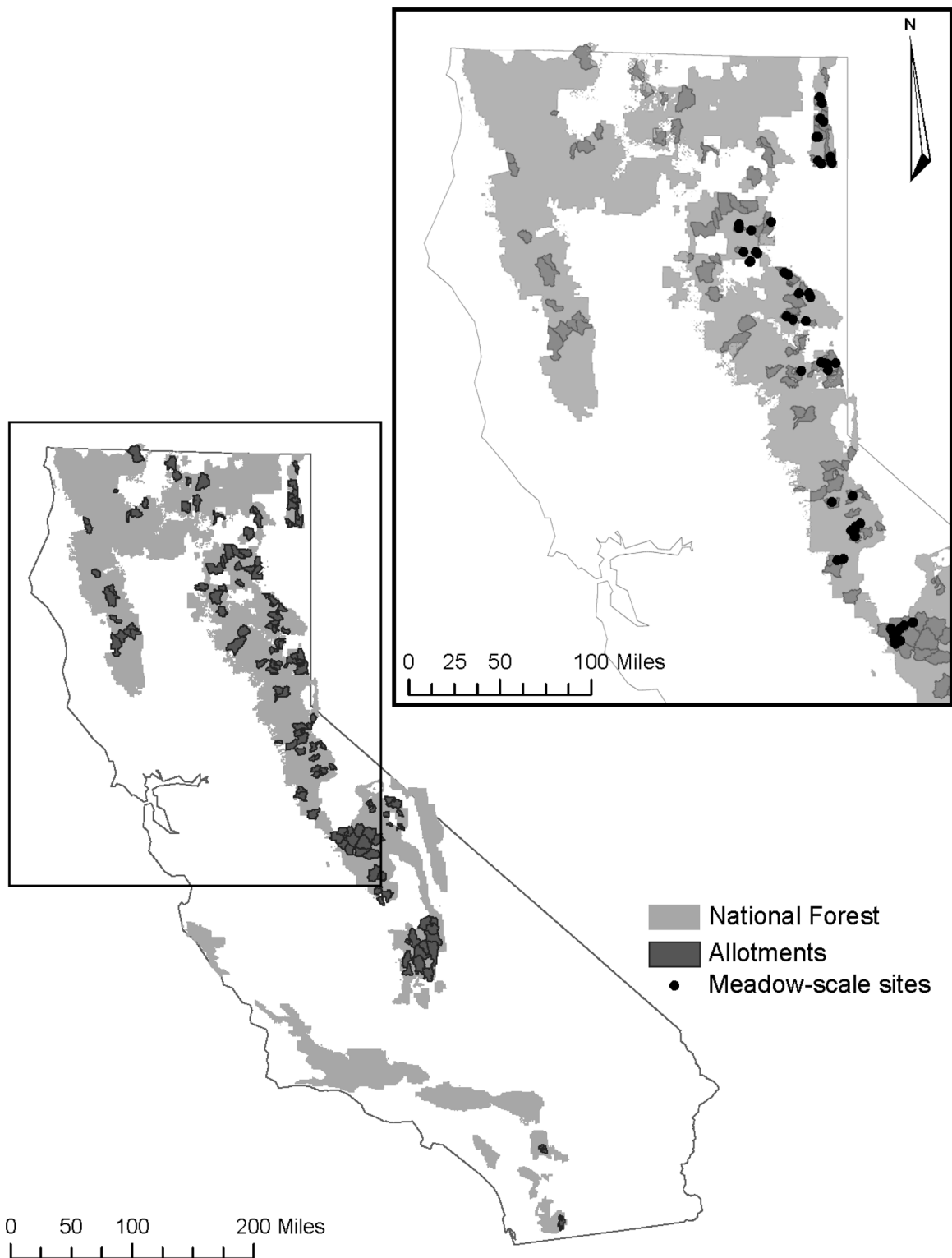


Fig. 2 Location of allotment and meadow study areas. *Light gray* shaded areas represent National Forest lands in California. *Dark gray* shaded areas represent the allotments in the allotment scale analysis of changes in plant community and each allotment contains at least one

long-term plant community monitoring site ($n = 279$). The *inset* shows a subset of monitoring sites that were selected for meadow scale analysis of changes plant community ($n = 52$)

Table 1 Summary of topographic, livestock grazing, and precipitation characteristics for sites at the allotment scale

	Elevation (m)	Latitude (dec. deg.)	Allotment area (km ²)	Stocking rate–10-year average (AUM/km ²) ^a	10-Year Cumulative (AUM/km ²) ^b	Total annual precipitation (30-year normal, mm)
Minimum	972	32.70	12	0.0	0	312
Median	1993	38.63	99	5.8	63	787
Mean	2027	38.63	121	7.9	83	906
Maximum	3257	42.06	311	35.8	365	2091

$n = 279$

AUM animal unit month

^a Average stocking rate during the 10 years preceding the final reading for each plant community monitoring plot, representing the average stocking rate during the plant community monitoring period.

^b Cumulative AUM during the plant community monitoring period standardized by allotment area.

Table 2 Summary of topographic, livestock grazing, and precipitation characteristics for sites at the meadow scale

	Elevation (m)	Latitude (dec. deg.)	Allotment area (km ²)	Stocking Rate–10-year average (AUM/km ²) ^a	10-year cumulative (AUM/km ²) ^b	Total annual precipitation (30-year normal, mm)
Minimum	1373	37.33	6	1.9	19	379
Median	1961	39.90	78	8.1	86	874
Mean	1943	39.49	83	11.6	117	948
Maximum	2490	41.67	181	33.7	337	1583

$n = 52$

AUM Animal unit month

^a Average stocking rate during the 10 years preceding the final reading for each plant community monitoring plot, representing the average stocking rate during the plant community monitoring period

^b Cumulative AUM during the plant community monitoring period standardized by allotment area

some grazing seasons began as early as February and ended as late as December. Across all 138 allotments used for allotment scale analysis, average annual allotment stocking rates ranged from 0 to 36 AUM/km² (Table 1). Stocking rates for the subset of 34 allotments used for meadow scale analysis were comparable to the full study set (Tables 1 and 2).

All grazing occurred under permits administered by the USFS, with riparian grazing standards established for meadow conservation. Specifically, three or more of the following riparian grazing standards have been in place on each grazed allotment since the initiation of plant community monitoring in 1997: (1) limits on herbaceous vegetation biomass consumption (e.g., <40% of annual production); (2) minimum allowable residual herbaceous vegetation height (e.g., >10 cm); (3) limits on browsing of recruiting riparian woody species (e.g., <20% of annual leader growth); and (4) limits on livestock hoof damage to soil and streambanks (e.g., <10% soil shearing by hooves) (Clary 1999; Clary and Leininger 2000; Clary and Webster 1990; Hall and Bryant 1995). The specific grazing management practices used in each allotment varied depending on site characteristics, but the riparian grazing standards were used as action triggers for grazing management at all study sites (US Forest Service 2004).

2.3 Data Collection

2.3.1 Plant community metrics

Herbaceous plant community monitoring sites were established between 1997 and 2005 at 279 riparian meadows. At each site, three parallel transects, 20 m long and 5 m apart, were established and permanently marked. Twenty 0.01 m² quadrats were positioned at 1 m intervals along each transect (60 total quadrats per site) during monitoring. Plant community composition data were collected by rooted species frequency methodology (Bonham 1989) at all sites approximately every 5 years between 1997 and 2013. At time of analysis, study monitoring plots had 2–7 readings (median = 3 readings), spanning 8 and 16 years of data collection (median = 10 years).

Rooted frequency data were used to calculate number of plant species (richness; S), Shannon Diversity Index (H'); and relative frequencies of forbs, non-native species, wetland obligate species, and upland species for each site and reading. S and H' were calculated using the vegan package (Oksanen et al. 2016) in the R statistical program (R Core Team 2015). Changes over time in plant community metrics were calculated as the difference between the first and last (i.e., most recent) readings for each monitoring plot

(henceforth indicated as Δ). The first and last readings vary between plots depending on when they were established and how often they are monitored.

2.3.2 Livestock grazing pressure–allotment scale

Livestock grazing pressure was calculated at the allotment scale as the cumulative annual stocking pressure (AUMs/km²) during the plant community monitoring period (10 years preceding the most recent reading) of each site. The year of the most recent reading (i.e., year 10) ranged from 2007 to 2013. An animal unit month (AUM) is the standard metric of grazing pressure for USFS managed grazing allotments. The cumulative stocking rate (10-year cumulative AUMs/km²) represents the full livestock grazing pressure on the allotment over the plant community monitoring period standardized by allotment area (km²). Annual authorized allotment stocking rates were compiled from USFS records.

2.3.3 Livestock grazing pressure–meadow scale

Because stocking rates are only assigned by the USFS at the allotment scale, meadow scale livestock grazing pressure was measured as livestock fecal density (pats/ha). Fecal density was measured at each of 52 plant community monitoring sites during the 2015 grazing season. Fecal density was measured at the end of the grazing season by counting all distinct fecal pats within three 20 m² belt transects (1.3 × 15.2 m) located on the plant community monitoring plots. Because livestock fecal pats persist up to 5–10 years in these sites, fecal density represents meadow scale cumulative livestock grazing pressure (Roche et al. 2012).

2.3.4 Environmental conditions–allotment scale

Relative precipitation was calculated as the mean percent of the 30-year normal precipitation (1981–2010; mm) allotments received annually (PRISM Climate Group 2015) during the monitoring period for each site within the allotment. First, both 30-year normal precipitation and total annual precipitation amounts were calculated for each allotment as the average of available values across the allotment area using gridded data from the PRISM Climate Group (PRISM Climate Group 2015). The PRISM Climate Group model uses weather station data (weighted by distance) and local geographical factors (e.g., elevation and aspect) to calculate a climate-elevation model for each 800 m grid cell (see Daly et al. 2008 for full description of PRISM methods). The 10-year mean total annual precipitation was calculated for the 10 years preceding the most recent reading for each plot within the allotment. The year

of the most recent reading (i.e., year 10) ranged from 2007 to 2013. Then, using allotment 30-year normal total annual precipitation values, relative precipitation was calculated as

$$\text{Relative precipitation} = \left(\frac{\text{10-year mean total annual precipitation}}{\text{30-year normal total annual precipitation}} \right) \times 100$$

Therefore, relative precipitation values below 100% indicate the 10-year period was drier than average for the respective allotments.

2.3.5 Environmental conditions–meadow scale

During the 2015 grazing season, site wetness was also recorded for the subset of 52 plant community monitoring sites. Site wetness was rated on a scale of 1–5, where 1 is the driest site and 5 is the wettest site (Roche et al. 2012; Roche et al. 2014) observed in this study. Plant community, soil characteristics, and hydrologic characteristics observed during data collection were all used to determine the site wetness rank for each site. For example, a site dominated by wetland obligate species on organic soils and visible standing water at the end of the grazing season would represent a 5 rank. A relatively drier grass/forb-dominated site on mineral soils would represent a 1 rank. Meadow scale relative precipitation was calculated following the methods outlined for allotment scale (Section 2.3.4) (PRISM Climate Group 2015). Meadow scale 30-year normal precipitation and total annual precipitation amounts were queried from the PRISM Climate Group (PRISM Climate Group 2015) for a single geographic point—the coordinates of the plant community monitoring site within each meadow—using the raster package (Hijmans 2015) in the R statistical program (R Core Team 2015).

2.4 Statistical Analysis

2.4.1 Allotment scale

Linear regression analysis was used to test for associations between changes (Δ) in plant community metrics at each monitoring site, and allotment scale livestock grazing pressure and environmental conditions during the study period (279 sites across 138 grazing allotments). Backward step-wise selection was used to determine final models, with $p < 0.1$ (using a Bonferroni correction of < 0.016 to account for six analyses) required for significance. The Bonferroni correction protects against Type I error when multiple comparisons are made by setting a lower threshold for significance. Allotment scale relationships with $p < 0.016$ are reported as “statistically significant”, while relationships with $p < 0.1$ are reported as “apparent”. Initial full model independent variables were: allotment elevation (m),

relative precipitation, cumulative stocking rate (AUM/km²), and the cumulative stocking rate by relative precipitation interaction term. Individual models were analyzed for six plant community response metrics (dependent variables): ΔS ; $\Delta H'$; and Δ relative frequency (%) of forbs, non-native species, wetland obligate species, and upland species. Regression analyses were conducted using the R statistical program (R Core Team 2015), and standard diagnostics were used to confirm that assumptions were met.

2.4.2 Meadow scale

Meadow scale analyses were conducted on a subset of 52 monitoring sites located across 34 grazing allotments (Fig. 2). In addition to the six meadow scale plant community change metrics described in Section 2.3.1, changes in overall meadow plant community composition during the study period were estimated via non-metric multi-dimensional scaling (NMDS) using a Euclidean distance matrix. Solutions were tested for 1 through 7 dimensions and a plot of final stress vs. number of dimensions was used to determine the optimal number of dimensions (McCune and Grace 2002). Composition was quantified using NMDS scores from the first dimension of the Euclidean distance matrix (the dimension that explains the most variation in the data) and changes in composition were calculated as the difference between the first and last years for each monitoring plot. NMDS analysis was completed using the metaMDS function in the vegan package (Oksanen et al. 2016) in the R statistical program (R Core Team 2015). Permutational multivariate analysis of variance (perMANOVA) with a Euclidean distance matrix was used to test for shifts in overall plant community composition between first and last years. This analysis was conducted using the adonis function in the vegan package of the R statistical program (Oksanen et al. 2016; R Core Team 2015).

Linear regression analysis was used to test for associations between changes (Δ) in plant community metrics at each monitoring site, and meadow (site) scale livestock grazing pressure and environmental conditions during the study period (52 sites within 34 grazing allotments). Backward step-wise selection was used to determine final models, with $p < 0.1$ (using a Bonferroni correction of < 0.014 to account for seven analyses) required for significance. Meadow scale relationships with $p < 0.014$ are reported as statistically significant, while relationships with $p < 0.1$ are reported as apparent. Initial full model independent variables were: meadow elevation (m), relative precipitation, wetness ranking, fecal density (pats/ha), and the fecal density by relative precipitation interaction term. Individual final models were analyzed for seven plant community response metrics (dependent variables): ΔS ; $\Delta H'$; Δ species composition (calculated from NMDS scores of

first and last reading years); and Δ relative frequency (%) of forbs, non-native species, wetland obligate species, and upland species. Regression analyses were conducted using the R statistical program (R Core Team 2015), and standard diagnostics were used to check conformity with assumptions.

3 Results

3.1 Allotment Scale

At the allotment scale, average precipitation during the 10 years preceding the most recent readings ranged from 75% to 110% of 30-year normal with a mean value of 94%. The allotment-scale median total annual precipitation during the study period was slightly lower than the median total annual precipitation during the preceding decade (data not shown). Median changes for non-native species and upland species were 0 (i.e., no change between first and last year) across all 138 allotments (279 monitoring sites) (Fig. 3). Median changes were positive for S, H', and forbs, while wetland obligate species median change was downward. Neither highly or moderately invasive species (California Invasive Plant Council 2016) were abundant with median relative frequencies and changes of 0. Moderately invasive species were observed in 22 allotments (30 of 279 sites), and highly invasive species were observed in 17 allotments (25 of 279 sites).

No significant (Bonferroni correction of $p < 0.016$) relationships were found between allotment scale livestock grazing pressure, relative allotment scale precipitation, or their interaction term with any of the plant community metrics examined (ΔS ; $\Delta H'$; and Δ relative frequency of forbs, non-native species, wetland obligate species, and upland species) (Figs. 4 and 5). There was an apparent, but not statistically significant, positive relationship ($p = 0.07$, slope = 0.28, adjusted $r^2 = 0.008$) between precipitation and changes in relative frequency of wetland obligate species (Fig. 5c). There was a significant negative relationship ($p = 0.009$, slope = -0.004 , adjusted $r^2 = 0.02$) between elevation and changes in relative frequency of upland species; and an apparent, but not statistically significant, negative relationship ($p = 0.07$, slope = -0.003 , adjusted $r^2 = 0.008$) between elevation and changes in relative frequency of non-native species (not shown). No other significant or apparent relationships were found.

3.2 Meadow Scale

At the meadow scale, average precipitation during the 10 years preceding the most recent reading ranged from 59% to 96% of 30-year normal with a mean value of 82%. The

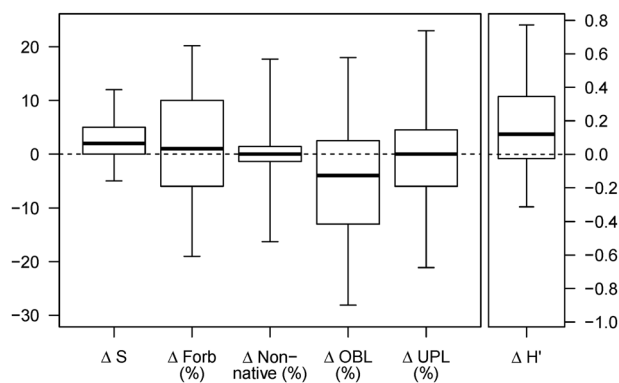


Fig. 3 Long-term change (Δ) in allotment scale plant community metrics in 279 plant community monitoring plots in riparian meadows across California ($n = 279$). Metrics are species richness (S), diversity (H'), and the relative frequencies of forb, non-native, wetland obligate (OBL), and upland species (UPL). Dark lines represent the median. Top and bottom box boundaries represent the 75th and 25th percentiles, respectively. Top and bottom whiskers represent the 95th and 5th percentiles, respectively

meadow-scale median total annual precipitation for the study period was similar to the median total annual precipitation during the decade prior to the study period (data not shown). The perMANOVA analysis indicated that plant community composition was significantly different between the first and last readings ($p < 0.001$), which indicates that plant communities in the study meadows have shifted in composition during the study period. Median changes across the subset of 52 monitoring sites for non-native species and upland species were 0 (i.e., no change between first and last year) (Fig. 6). Median changes in S, H' , and forbs were positive, while changes in wetland obligate species was negative. Neither highly or moderately invasive species (California Invasive Plant Council 2016) were abundant with median relative frequencies and changes of 0.

Significant (Bonferroni correction of $p < 0.014$) relationships were found for three of the seven plant community metrics analyzed (Table 3). However, adjusted r^2 values (0.18 to 0.25) indicate that, while these regression models were significant, substantial variation was left unexplained by meadow scale livestock grazing pressure and environmental conditions observed during the study period. Relative precipitation was significantly related to changes in forbs ($p < 0.001$) and wetland obligate species ($p = 0.01$), as well as apparently related to changes in non-native species ($p = 0.06$) (Fig. 7, Table 3). Fecal density was significantly and negatively related to changes in wetland obligate species ($p = 0.01$) and upland species ($p = 0.01$) (Fig. 8b and c, Table 3). A significant ($p = 0.01$) fecal density by relative precipitation interaction indicated a more positive association between fecal density and change in upland species for sites experiencing drier conditions (Table 3). Fecal density was apparently ($p = 0.09$)

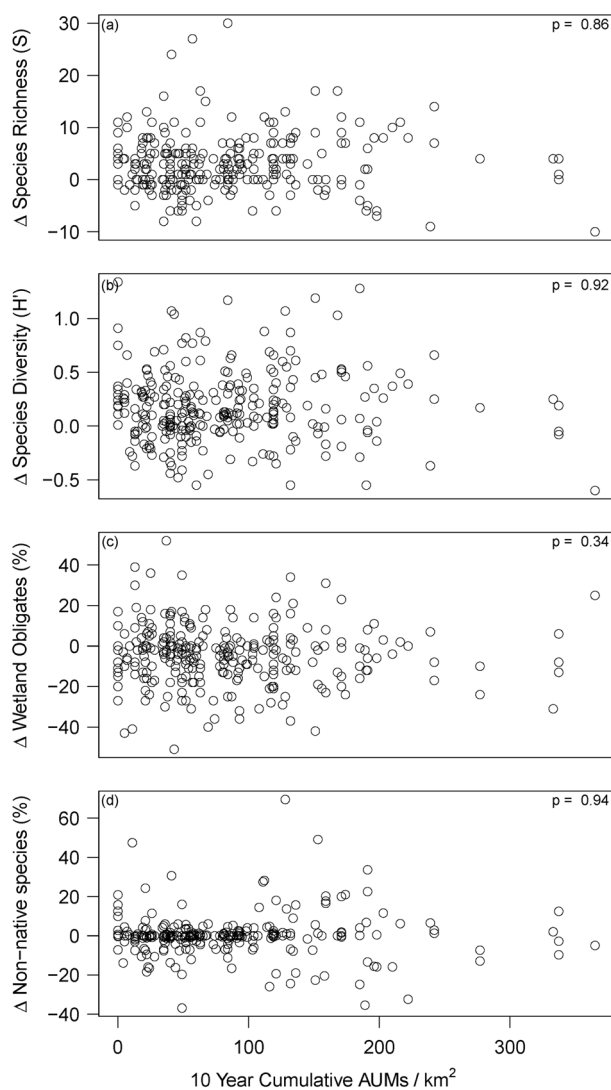


Fig. 4 Allotment scale changes (Δ) in (a) species richness (S), (b) species diversity (H'), and the relative frequencies of (c) wetland obligate and (d) non-native species (%) by the 10-year cumulative allotment stocking rate (AUM/km²; $n = 279$). As indicated by the p values listed in each panel, no significant relationships were found between these plant community metrics and cumulative stocking rate at the allotment scale (Bonferroni-corrected $p < 0.016$)

negatively related to changes in species richness (Fig. 8a). Meadow site wetness rank (1 through 5) was apparently ($p = 0.07$) associated with changes in meadow plant community composition (calculated from NMDS scores of first and last reading years) (Table 3). Plant community composition scores (via NMDS) are reported in Online Resource 1.

4 Discussion

Plant community composition across the 279 riparian meadows enrolled in this study has been changing over the

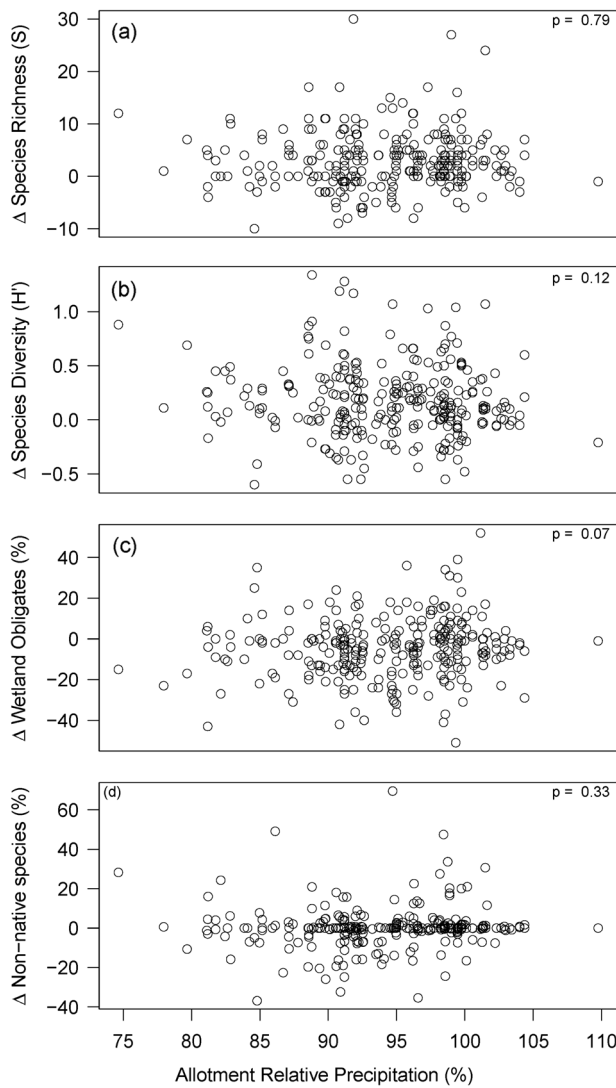


Fig. 5 Allotment scale changes (Δ) in (a) species richness (S), (b) species diversity (H'), and the relative frequencies of (c) wetland obligate and (d) non-native species (%) by the 10-year average relative precipitation (percent of 30-year normal annual precipitation; $n = 279$). As indicated by the p values listed in each panel, no significant relationships were found between these plant community metrics and cumulative stocking rate at the allotment scale (Bonferroni-corrected $p < 0.014$). One apparent relationship (i.e., $p < 0.1$) was found between changes in wetland obligate species and relative precipitation

past decade, with substantial variation in magnitude and direction of change found between meadows (Fig. 3). Overall, median changes in plant community metrics suggest that species diversity and richness are increasing, and that the non-native and upland species components of these meadow communities have been constant. Invasive plant species frequency has remained at low levels throughout this period, with increases in overall species richness and diversity driven by increases in native forbs. These outcomes are congruent with meadow conservation objectives. The largest observed functional change was a reduction in

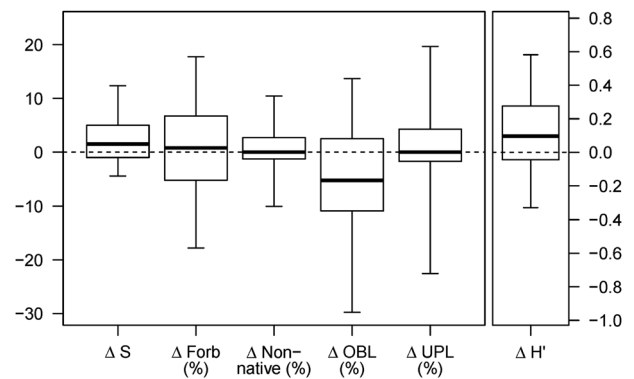


Fig. 6 Long-term trends (Δ) in meadow scale plant community metrics in 52 grazed meadows. Richness (S), diversity (H'), and the relative frequencies of forb, non-native, wetland obligate (OBL), and upland (UPL) species. Dark lines represent the median. Top and bottom box boundaries represent the 75th and 25th percentiles, respectively. Top and bottom whiskers represent the 95th and 5th percentiles, respectively

the relative frequency of wetland obligate species during the study period. The majority of the meadows studied received precipitation below the 30-year normal throughout the study decade.

Associations between livestock grazing pressure and changes in plant community varied across spatial scales. Our results suggest that allotment scale livestock grazing pressure (i.e., stocking rate) is currently at a level that balances production and conservation goals at the allotment scale. During the study period, total grazing pressure declined 36% across USFS administered grazing allotments in California, second only to Wyoming in the 11 western states (Fig. 1). The reduced allotment scale grazing pressure observed in this study was not associated with changes in plant community metrics (Fig. 4). However, there were slight negative associations between meadow scale grazing pressure and changes in three of the seven metrics examined (Fig. 8, Table 3). This suggests that at the current reduced stocking rates, heterogeneity of livestock use within an allotment is a significant driver of changes in meadow plant community. Meadow scale results confirm the importance of adaptively managing annual grazing pressure on individual meadows within allotments (Clary 1999; Freitas et al. 2014; George et al. 2011) even when allotment stocking rates are relatively low. Grazing management and distribution strategies (e.g., off stream water development) can be effective in limiting cattle use of sensitive habitats like riparian areas (Johnson et al. 2016). Implementation and monitoring of livestock grazing distribution strategies to meet riparian grazing standards in all meadows within allotments is key to obtaining conservation objectives.

Meadow scale monitoring is especially important to adjust grazing management as climate conditions change over time. Precipitation timing and frequency is expected to

Table 3 Meadow-scale analysis of relationships between plant community trends and management and environmental factors on USFS lands in California

Response	Parameter	Coefficient	P-value
ΔS (richness) (adj. $r^2 = 0.04$)	Elevation	<i>ns</i>	0.56
	Site wetness rank	<i>ns</i>	0.80
	Precipitation percent of normal	<i>ns</i>	0.37
	Fecal pats	1.32	0.09
	Precipitation * fecal Pats	<i>ns</i>	0.61
$\Delta H'$ (diversity)	Elevation	<i>ns</i>	0.37
	Site wetness rank	<i>ns</i>	0.45
	Precipitation percent of normal	<i>ns</i>	0.16
	Fecal pats	<i>ns</i>	0.63
	Precipitation * fecal pats	<i>ns</i>	0.70
ΔForbs (%) (adj. $r^2 = 0.22$)	Elevation	<i>ns</i>	0.34
	Site wetness rank	<i>ns</i>	0.71
	Precipitation percent of normal	-5.737	< 0.001*
	Fecal pats	<i>ns</i>	0.37
	Precipitation * fecal pats	<i>ns</i>	0.10
$\Delta \text{Non-native species}$ (%) (adj. $r^2 = 0.05$)	Elevation	<i>ns</i>	0.37
	Site wetness rank	<i>ns</i>	0.95
	Precipitation percent of normal	2.173	0.06
	Fecal pats	<i>ns</i>	0.39
	Precipitation * fecal pats	<i>ns</i>	0.79
$\Delta \text{Wetland obligate species}$ (%) (adj. $r^2 = 0.25$)	Elevation	<i>ns</i>	0.65
	Site wetness rank	<i>ns</i>	0.50
	Precipitation percent of normal	4.458	0.01*
	Fecal pats	-4.260	0.01*
	Precipitation * fecal pats	<i>ns</i>	0.35
$\Delta \text{Upland species}$ (%) (adj. $r^2 = 0.18$)	Elevation	<i>ns</i>	0.89
	Site wetness rank	<i>ns</i>	0.81
	Precipitation percent of normal	-4.604	0.32
	Fecal pats	-2.732	0.01*
	Precipitation * fecal pats	-6.2772	0.01*
$\Delta \text{Species composition}$ (adj. $r^2 = 0.05$)	Elevation	<i>ns</i>	0.39
	Site wetness rank	1.257	0.07
	Precipitation percent of normal	<i>ns</i>	0.64
	Fecal pats	<i>ns</i>	0.36
	Precipitation * fecal pats	<i>ns</i>	0.21

$n = 52$

Δ indicates the trend in a given plant community metric during the monitoring period. Final model coefficients are centered on the mean and scaled to one standard deviation. Coefficients are given for parameters with apparent relationships with responses ($p < 0.1$). "ns" indicates no apparent relationship between the parameter and response ($p \geq 0.1$) and the reported p -value is the final p -value for that parameter before it was dropped from the model

*indicates the parameter was significant using Bonferroni corrected $p < 0.016$

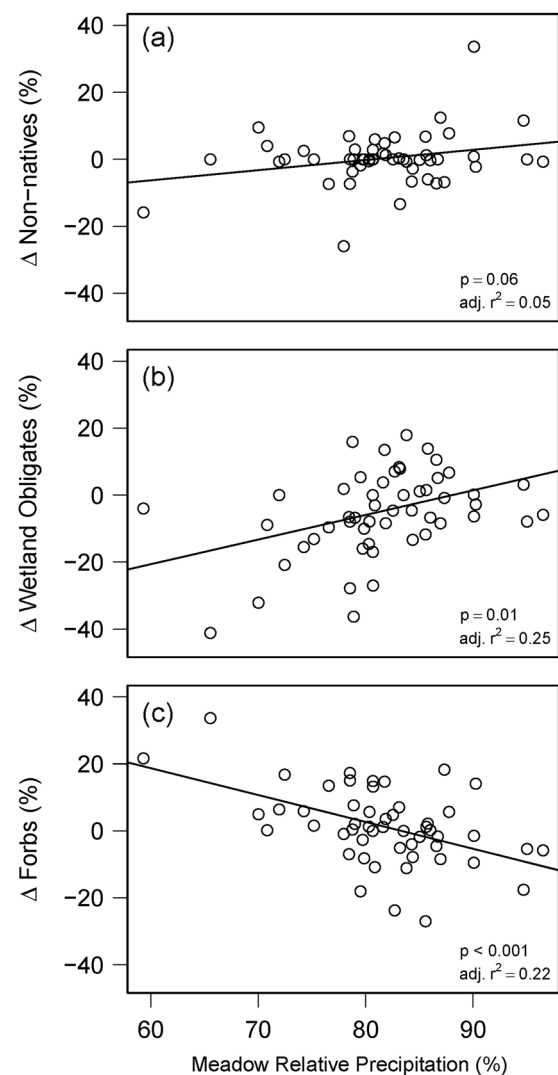


Fig. 7 Meadow scale changes (Δ) in the relative frequencies (%) of (a) non-native species, (b) wetland obligate species, and (c) forbs by precipitation as a percent of long-term normal levels ($n = 52$). The effect of precipitation was positive for wetland obligate species and negative for forbs (Bonferroni-corrected $p < 0.016$). There was an apparent (i.e., $p < 0.1$) positive effect of precipitation for non-native species. P values are presented for the effect of meadow relative precipitation (10-year average annual precipitation percent of 30-year normal annual precipitation) in each model

become more unpredictable, especially at higher elevations (Cayan et al. 2008; Maurer 2007). Most studies predict that these precipitation changes will lead to earlier spring runoff, reduced annual snowpack and stream flow, and more winter rainfall-runoff—resulting in a longer and drier summer growing season (Miller et al. 2003; Null et al. 2010; Mote et al. 2005; Stewart et al. 2005). These changes would likely result in reduced extent and duration of soil saturation and anaerobic conditions in wetter meadows. Results from this study (Table 3) and previous studies (Allen-Diaz 1991; Dwire et al. 2004; Freitas et al. 2014) show that plant

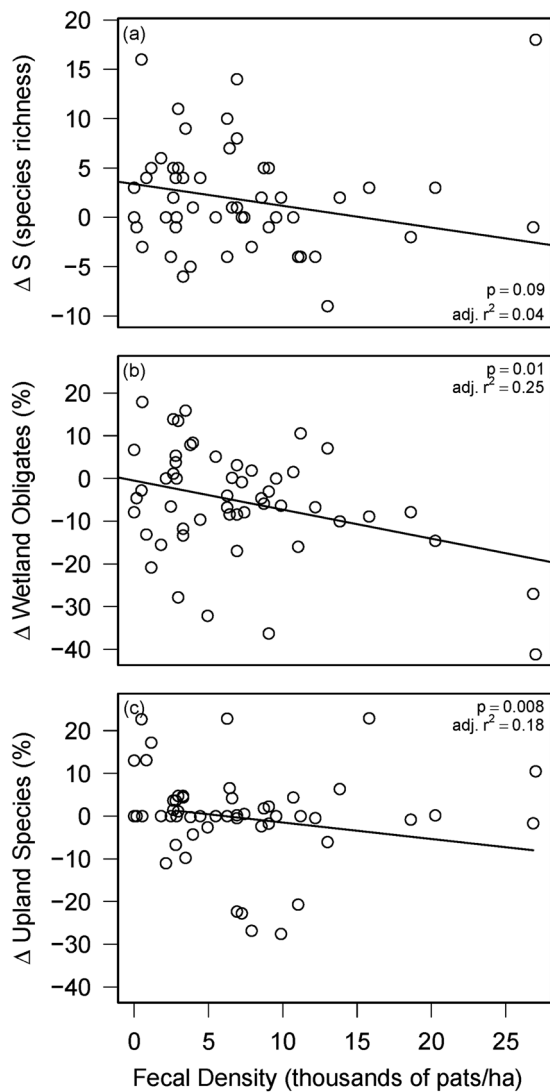


Fig. 8 Meadow scale changes (Δ) in species richness (S) and the relative frequency (%) of wetland obligate species and upland species by fecal density ($n=52$). Fecal density had a significant negative effect of on changes in wetland obligate species (Bonferroni-corrected $p < 0.016$). The fecal density by relative precipitation interaction was significant for changes in upland species. There was an apparent (i.e., $p < 0.1$) negative effect of fecal density on changes in species richness. In panels **a** and **b**, p values are presented for the fecal density effect. In panel **c**, the p value shown is for the interaction of fecal density and relative precipitation and the line represents the effect of fecal density at the median relative precipitation value (81.7% of 30-year normal annual precipitation). The line has been clipped to the range of fecal densities observed at sites receiving approximately 81.7% relative precipitation

community composition in these meadows is strongly linked to hydrological processes (e.g., site wetness and precipitation). Shifts in plant community composition associated with hydrological processes are important to consider in grazing management decisions as climate patterns become more variable, even for plant community metrics that were not associated with grazing pressure (e.g., forbs and non-native

species, Table 3). Changes in precipitation amount and timing, as well as any associated effects on stream flow and water table dynamics, will likely lead to changes in plant community structure and, consequently, the timing, amount, and quality of forage available for livestock grazing (Izaurre et al. 2011; Null et al. 2010). Additionally, grazing pressure interacted with precipitation for a plant community metric at the meadow scale (Δ upland species, Table 3), suggesting that plant community responses to grazing could also become more varied as climate changes. Because individual meadows are likely to experience and respond to changing climate conditions in site-specific ways, site-level monitoring and grazing management will be important to assess meadow trends and response to grazing.

Since the mid-20th century, public grazing lands policy and management paradigms have moved to increasingly integrate agricultural and conservation goals. In the modern public lands grazing era, riparian meadows are part of a working landscape that is managed to balance a multitude of economic, social, and ecological goals. Our work, and the work of others (Clary and Leininger 2000; Clary and Webster 1990; Freitas et al. 2014; Hall and Bryant 1995), suggests that current riparian grazing utilization limits can effectively protect these meadow resources. Adaptive implementation of site-specific riparian grazing conservation strategies to meet these limits is critical to safeguard ecosystem services, particularly under increasingly variable and changing environmental conditions.

Acknowledgements This research was funded by the U.S. Forest Service Pacific Southwest Region, Russell L. Rustici Research and Cattle Endowment, and the UC Davis Department of Plant Sciences Graduate Student Research Assistantship Program. Anton Jackson, Anne Yost, Tom Froli, Kendra Brown, Neil McDougald, Tracy Schorr, DJ Eastburn, and USFS ranger district staff members assisted in developing this project, compiling grazing records, and collecting monitoring data.

Compliance with ethical standards

Conflict of interest The authors declare that they have no competing interests.

Open Access This article is distributed under the terms of the Creative Commons Attribution 4.0 International License (<http://creativecommons.org/licenses/by/4.0/>), which permits unrestricted use, distribution, and reproduction in any medium, provided you give appropriate credit to the original author(s) and the source, provide a link to the Creative Commons license, and indicate if changes were made.

References

Acreman M, Holden J (2013) Hows wetlands affect floods. *Wetlands* 33:773–786. doi:10.1007/s13157-013-0473-2

- Allen-Diaz BH (1991) Water table and plant species relationships in Sierra Nevada meadows. *Am Midl Nat* 126:30–43
- Armour CL, Duff DA, Elmore W (1991) The Effects of Livestock Grazing on Riparian and Stream Ecosystems. *Fisheries* 16:7–11
- Bales RC et al. (2011) Forests and Water in the Sierra Nevada Watershed Ecosystem Enhancement Project. Sierra Nevada Research Institute. Merced, CA. Report Number 11.1
- Belsky AJ, Matzke A, Uselman S (1999) Survey of livestock influences on stream and riparian ecosystems in the western United States. *J Soil Water Conserv* 54:419–431
- Bonham CD (1989) Measurements for terrestrial vegetation. Wiley, New York, NY
- Briske, DD (ed) (2011) Conservation benefits of rangeland practices: Assessment, recommendations, and knowledge gaps. U.S. Dept. of Agriculture, Natural Resources Conservation Service, Allen Press, Inc., Lawrence, KS
- Bureau of Land Management (2016) Public Land Statistics. http://www.blm.gov/public_land_statistics/. Accessed 7 Sep 2016
- California Invasive Plant Council (2016) California invasive plant inventory database. <http://www.cal-ipc.org/paf/>. Accessed 1 May 2015
- Cayan DR, Maurer EP, Dettinger MD, Tyree M, Hayhoe K (2008) Climate change scenarios for the California region. *Clim Change* 87:S21–S42. doi:10.1007/s10584-007-9377-6
- Clary WP (1999) Stream channel and vegetation responses to late spring cattle grazing. *J Range Manage* 52:218–227
- Clary WP, Leininger WC (2000) Stubble height as a tool for management of riparian areas. *J Range Manage* 53:562–573. doi:10.2307/4003148
- Clary WP, Webster BF (1990) Riparian grazing guidelines for the intermountain region. *Rangelands* 12:4
- Daly C, Halbleib M, Smith JI, Gibson WP, Doggett MK, Taylor GH, Curtis J, Pasteris PP (2008) Physiographically sensitive mapping of climatological temperature and precipitation across the conterminous United States. *Int J Climatol*. doi:10.1002/joc.1688
- Davy JS, Roche LM, Robertson AV, Nay DE, Tate KW (2015) Introducing cattle grazing to a noxious weed-dominated rangeland shifts plant communities. *Calif Agric* 69:230–236. doi:10.3733/ca.v069n04p230
- Dwire KA, Kauffman JB, Brookshire EN, Baham JE (2004) Plant biomass and species composition along an environmental gradient in montane riparian meadows. *Oecologia* 139:309–317. doi:10.1007/s00442-004-1498-2
- Eldridge DJ, Greene RSB (1994) Microbiotic soil crusts - a review of their roles in soil and ecological processes in the rangelands of Australia. *Aust J Soil Res* 32:389–415. doi:10.1071/Sr9940389
- Fleischner TL (1994) Ecological costs of livestock grazing in western north America. *Conserv Biol* 8:629–644
- Freitas MR, Roche LM, Weixelman D, Tate KW (2014) Montane meadow plant community response to livestock grazing. *Environ Manage* 54:301–308. doi:10.1007/s00267-014-0294-y
- George MR, Jackson RD, Boyd CS, Tate KW (2011) A scientific assessment of the effectiveness of riparian management practices. In: Briske DD (ed) Conservation benefits of rangeland practices: assessment, recommendations, and knowledge gaps. U.S. Dept. of Agriculture, Natural Resources Conservation Service, Allen Press, Inc., Lawrence, KS, pp 213–252
- Hall FC, Bryant L (1995) Herbaceous stubble height as a warning of impending cattle grazing damage to riparian areas. U.S. Department of Agriculture, Forest Service, Pacific Northwest Research Station, Portland, OR. Gen. Tech. Rep. PNW-GTR-362
- Hammersmark CT, Dobrowski SZ, Rains MC, Mount JF (2010) Simulated effects of stream restoration on the distribution of wet-meadow vegetation. *Restor Ecol* 18:882–893. doi:10.1111/j.1526-100X.2009.00519.x
- Hatfield RG, LeBuhn G (2007) Patch and landscape factors shape community assemblage of bumble bees, *Bombus* spp. (Hymenoptera: Apidae), in montane meadows. *Biol Conserv* 139:150–158. doi:10.1016/j.biocon.2007.06.019
- Hijmans RJ (2015) Raster: Geographic data analysis and modeling. R package version 2.5-8. <http://CRAN.R-project.org/package=raster>
- Hormay AL, Evanko AB (1958) Rest-rotation grazing: a management system for bunchgrass ranges. California Forest and Range Experiment Station, Albany, CA
- Izaurrealde RC, Thomson AM, Morgan JA, Fay PA, Polley HW, Hatfield JL (2011) Climate impacts on agriculture: implications for forage and rangeland production. *Agron J* 103:371–381. doi:10.2134/agronj2010.0304
- Johnson D, Larson L, Wilson K, Clark P, Williams J, Louhaichi M (2016) Cattle use of perennial streams and associated riparian areas on a northeastern Oregon landscape. *J Soil Water Conserv* 71:484–493
- Kauffman JB, Krueger WC (1984) Livestock impacts on riparian ecosystems and streamside management implications - a Review. *J Range Manage* 37:430–438. doi:10.2307/3899631
- Kleinfelder D, Swanson S, Norris G, Clary W (1992) Unconfined compressive strength of some streambank soils with herbaceous roots. *Soil Sci Soc Am J* 56:1920–1925
- Kuhn TJ, Safford HD, Jones BE, Tate KW (2011) Aspen (*Populus tremuloides*) stands and their contribution to plant diversity in a semiarid coniferous landscape. *Plant Ecol* 212:1451–1463. doi:10.1007/s11258-011-9920-4
- Manning ME, Swanson SR, Svejcar T, Trent J (1989) Rooting characteristics of four intermountain meadow community types. *J Range Manage* 42:309–312. doi:10.2307/3899500
- Matejkova I, van Diggelen R, Prach K (2003) An attempt to restore a central European species-rich mountain grassland through grazing. *Appl Veg Sci* 6:161–168
- Maurer EP (2007) Uncertainty in hydrologic impacts of climate change in the Sierra Nevada, California, under two emissions scenarios. *Clim Change* 82:309–325. doi:10.1007/s10584-006-9180-9
- McCune B, Grace JB (2002) Analysis of Ecological Communities. MjM Software Design, Gleneden Beach, OR
- Meehan WR, Platts WS (1978) Livestock grazing and the aquatic environment. *J Soil Water Conserv* 33:274–278
- Micheli ER, Kirchner JW (2002) Effects of wet meadow riparian vegetation on streambank erosion. 2. Measurements of vegetated bank strength and consequences for failure mechanics. *Earth Surf Process Land* 27:687–697. doi:10.1002/esp.340
- Middleton BA, Holsten B, van Diggelen R (2006) Biodiversity management of fens and fen meadows by grazing, cutting and burning. *Appl Veg Sci* 9:307–316. doi:10.1658/1402-2001(2006)9[307:Bmofaf]2.0.Co;2
- Miller NL, Bashford KE, Strem E (2003) Potential impacts of climate change on California hydrology. *J Am Water Resour Assoc* 39:771–784. doi:10.1111/j.1752-1688.2003.tb04404.x
- Mote PW, Hamlet AF, Clark MP, Lettenmaier DP (2005) Declining mountain snowpack in Western North America. *Bull Am Meteorol Soc*. doi:10.1175/BAMS-86-1-39
- Norton JB, Jungst LJ, Norton U, Olsen HR, Tate KW, Horwath WR (2011) Soil carbon and nitrogen storage in upper montane riparian meadows. *Ecosystems* 14:1217–1231. doi:10.1007/s10021-011-9477-z
- Null SE, Viers JH, Mount JF (2010) Hydrologic response and watershed sensitivity to climate warming in California's sierra nevada. *Plos One* 5:ARTN e9932. doi:10.1371/journal.pone.0009932
- Odion DC, Dudley TL, D'Antonio CM (1988) Cattle grazing in southeastern Sierran meadows: ecosystem change and prospects

- for recovery. In: Hall CAJ, Doyle-Jones V (eds) Plant biology of Eastern California natural history of the White-Inyo range symposium, vol 2. University of California White Mountain Research Station, Los Angeles, CA, pp 277–292
- Oksanen J, Blanchet FG, Friendly M, et al. (2016) vegan: Community Ecology Package, R package version 2.3-3. <http://CRAN.R-project.org/package=vegan>
- PRISM Climate Group (2015) Oregon State University. <http://www.prism.oregonstate.edu/>. Accessed 27 Oct 2015
- Pyke CR, Marty J (2005) Cattle grazing mediates climate change impacts on ephemeral wetlands. *Conserv Biol* 19:1619–1625. doi:10.1111/j.1523-1739.2005.00233.x
- R Core Team (2015) R: A language and environment for statistical computing. R Foundation for Statistical Computing, Vienna. <https://www.R-project.org/>
- Ratliff RD, Reppert JN, McConnen RJ (1972) Rest-rotation grazing at Harvey Valley... range health, cattle gains, costs. U.S. Department of Agriculture, Forest Service, Berkeley, CA. USDA Forest Service Research Paper PSW - 77 /1972
- Roche LM, Latimer AM, Eastburn DJ, Tate KW (2012) Cattle grazing and conservation of a meadow-dependent amphibian species in the Sierra Nevada. *PLoS One* 7:e35734. doi:10.1371/journal.pone.0035734
- Roche LM, O'Geen AT, Latimer AM, Eastburn DJ (2014) Montane meadow hydroperiodology, plant community, and herbivore dynamics. *Ecosphere* 5:art150. doi:10.1890/es14-00173.1
- Rosenthal G, Schrautzer J, Eichberg C (2012) Low-intensity grazing with domestic herbivores: A tool for maintaining and restoring plant diversity in temperate Europe. *Tuexenia* 32:167–205
- Sampson AW, Weyl LH (1918) Range preservation and its relation to erosion control on western grazing lands. US Department of Agriculture, Washington, DC, Bulletin No. 675
- Stewart IT, Cayan DR, Dettinger MD (2005) Changes toward earlier streamflow timing across Western North America. *J Clim* 18:1136–1155. doi:10.1175/JCLI3321.1
- U.S. Government Accounting Office (1988) Public rangelands: some riparian areas restored but widespread improvement will be slow. U.S. Government Accounting Office, Washington, D.C. GAO/RCED-88-105
- US Forest Service (2004) Sierra Nevada Forest Plan Amendment: Final Supplemental Environmental Impact Statement, vol 1. US Department of Agriculture, Forest Service, Pacific Southwest Region, Vallejo, CA. R5-MB-046
- US Forest Service (2016) Grazing Statistical Summary Reports. <http://www.fs.fed.us/rangeland-management/reports/index.shtml>. Accessed Sep 7 2016