Chapter 5: California Spotted Owls

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Introduction

California spotted owls (Strix occidentalis occidentalis) are habitat specialists that are strongly associated with late-successional forests. For nesting and roosting, they require large trees and snags embedded in a stand with a complex forest structure (Blakesley et al. 2005, Gutiérrez et al. 1992, Verner et al. 1992b). In mixed-conifer forests of the Sierra Nevada, California spotted owls typically nest and roost in stands with high canopy closure (≥75 percent) [Note: when citing studies, we use terminology consistent with Jennings et al. (1999), however, not all studies properly distinguish between canopy cover and closure and often use the terms interchangeably (see chapter 14 for clarification)] and an abundance of large trees (>24 in (60 cm) diameter at breast height [d.b.h.]) (Bias and Gutiérrez 1992, Gutiérrez et al. 1992, LaHaye et al. 1997, Moen and Gutiérrez 1997, Verner et al. 1992a). The California spotted owl guidelines (Verner et al. 1992b) effectively summarized much of the information about nesting and roosting habitat. Since that report, research on the California spotted owl has continued with much of the new information concentrated in five areas: population trends, barred owl (Strix varia) invasion, climate effects, foraging habitat, and owl response to fire.

Population Trends

A rangewide investigation from 1990 to 2005 into the population dynamics of the California spotted owl showed that subpopulations at four studied locations were declining or remaining steady (mean λ = 1.007, 95 percent confidence interval (CI) = 0.952 to

Summary of Findings

1. **Spotted owls select habitat at multiple scales**, with less flexibility in the nesting and roosting habitat requirements, and more flexibility in the foraging habitat.

2. **Foraging habitat appears to have more moderate canopy closure and is still associated with large trees**, possibly because of their importance as nest sites for northern flying squirrels, an important prey species for spotted owls in mesic Sierra Nevada forests.

3. **Low- to moderate-severity fire does not reduce the probability of spotted owl occupancy** if numerous large trees and areas of high canopy closure remain after a fire.

4. **Dense understory of regenerating trees can interfere with owl foraging.** Low- to moderate-severity fire reduces the density of small trees and may improve the habitat quality of spotted owl nesting or foraging habitat.

5. **Forest heterogeneity, with various vegetation communities or fire severities infused into late-successional forest, may improve spotted owl fitness.**

6. **Fire effects on foraging habitat are not well understood**, and future research needs to be directed toward owl foraging use patterns in a burned landscape.

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1.066) (Blakesley et al. 2010). Apparent survival was similar between the sexes and increased with owl age. The subpopulation residing in the only national park included in the meta-analysis, Sequoia-Kings Canyon (SEKI), showed the highest survival rates. Mean annual reproductive output (number of young fledged per territorial female) ranged from 0.988 (± 0.154) in the El Dorado National Forest to 0.555 (± 0.110) in SEKI. El Dorado showed the highest annual variation, with higher reproduction every second year, while SEKI had low annual variation. Although reproductive output varied between the four subpopulations, the El Dorado showed a declining trend. This declining trend is probably related to recent low annual reproduction and a consistent decrease in recruitment in the El Dorado. The high annual variation and decreasing trend observed in the El Dorado subpopulation indicate that habitat quality is not stable and probably decreasing over time. As with the northern spotted owl, reproductive output was highest with adults, followed by second-year subadults, and then first-year subadults. Population viability analyses indicated that the probability of a >10 percent decline in 7 years was lowest at 0.41 (95 percent CI = 0.09 to 0.78) for the Sierra National Forest subpopulation and highest at 0.64 (95 percent CI= 0.27 to 0.94) for the Lassen National Forest, and inconclusive for the El Dorado and SEKI (Blakesley et al. 2010).

**Barred Owl**

The barred owl, an aggressive competitor, has invaded the Sierra Nevada from the north and started reproducing as far south as the El Dorado National Forest (Keane 2007). This invasion appears to be a natural biogeographical process (Dark et al. 1998). Once barred owls establish a population adjacent to spotted owls, there are negative effects on spotted owl metapopulation dynamics (Olson et al. 2005). Barred owls are habitat and diet generalists (Livezey 2007) and appear to outcompete spotted owls during the breeding season, displacing spotted owls from territories that they occupy (Hamer et al. 2007). Ishak et al. (2008) found that many spotted owls had a blood parasite not found in barred owls, which may further reduce spotted owl competitiveness with the recent invader. In the Pacific Northwest, scientists and managers are currently trying to rapidly formulate methods for reducing the barred owls’ negative impacts on spotted owls (Buchanan et al. 2007).

**Climate Change Effects**

In the Pacific Northwest, Glenn et al. (2010) showed that the rate of population change (λ) for the northern spotted owl was positively influenced by wetter-than-normal growing seasons, which they speculated improved owl prey availability. However, cold, wet winter and spring (early nesting season), as well as the number
of hot summer days, negatively affected the rate of population change. The influence of climate data on the rate of population change (\( \lambda \)) was highly variable, explaining 3 to 85 percent of the total variability seen in \( \lambda \), across six different study areas. Adult survival, which was closely related to regional climate conditions, had a stronger influence on the rate of population change than recruitment, which was associated with local weather. As climate change models project warmer winters with higher variability in winter precipitation, and hotter, drier summers across the Pacific Northwest and northern California, climate could potentially have a rangewide negative effect on spotted owl survival, recruitment, and population growth rates. North et al. (2000) suggested that regional weather during the nesting season influenced reproductive success and nest-site canopy structure was important in mitigating the effects of detrimental weather. Carroll (2010) advised that models used for spotted owl conservation planning should incorporate habitat variables along with climate information.

**Spotted Owl Nesting and Foraging Habitat Characteristics**

Generally, spotted owl survival increases with increasing area of late-successional forest (Dugger et al. 2005, Franklin et al. 2000, Olson et al. 2004) and decreases with increasing area of early successional forests (Dugger et al. 2005). However, because owls use a variety of habitats for foraging and nesting, forest heterogeneity across the landscape can improve spotted owl viability. Spotted owl survival and reproductive rates were higher in owl territories that included a mosaic of vegetation types infused within late-successional forest (Franklin et al. 2000), presumably because there was a greater diversity or abundance of prey within this mosaic (Ward et al. 1998, Zabel et al. 1995).

Spotted owls select habitat at multiple spatial and temporal scales, with less flexibility in nesting and roosting habitat requirements than foraging habitat. Blakesley et al. (2005) used remote sensing vegetation data to investigate the importance of spatial scale for spotted owl occupancy on a landscape scale. Between the two scales (500 and 2,000 ac) (202 to 809 ha) they found that the forest structure at the 500-ac (202-ha) scale was the most important. Within that scale, studies agree that both high overstory canopy closure and cover and an abundance of large trees are major influences in owl habitat selection (Bias and Gutiérrez 1992, Blakesley et al. 2005, Gutiérrez et al. 1992, LaHaye et al. 1997, Moen and Gutiérrez 1997, Roberts et al. 2011, Verner et al. 1992a). Reproduction can be associated with foraging habitat quality, because owls appear to fledge young more often as prey availability increases (Carey et al. 1992, Rosenberg et al. 2003). In northeastern California,
spotted owl reproduction was negatively correlated to nonforested areas and forest types not used for nesting or foraging within the nest area (500 acres) (Blakesley et al. 2005).

While foraging has been better studied for the northern spotted owl (*Strix occidentalis caurina* ex. Forsman et al. 2004), there are several emerging patterns applicable to California spotted owl foraging. For example, northern spotted owls may forage in or near edge habitat (Clark 2007, Folliard et al. 2000, Ward et al. 1998), but California spotted owls did not locate their nests close to edges (Phillips et al. 2010). For California spotted owls, foraging habitat appears to be more open (≥40 percent) than nesting habitat (≥70 percent) with respect to canopy closure (Call et al. 1992, Zabel et al. 1992), basal area (Roberts et al. 2011), and stand density (Irwin et al. 2007). Additionally, spotted owl foraging habitat is associated with large trees, possibly because of their importance as nest sites for northern flying squirrels (*Glaucomys sabrinus*) (Meyer et al. 2005, Waters and Zabel 1995), an important prey species for spotted owls in mesic or closed-canopied Sierra Nevada forests (Williams et al. 1992). Irwin et al. (2007) found owl foraging associated with forests in proximity to nest sites and small streams. In an analysis of owl locations including many foraging locations in the southern Sierra Nevada, owls used canyon/stream bottoms significantly more than expected (Underwood et al. 2010). Riparian area use may be related to preferred forest structural conditions (i.e., large trees and high canopy closure) or possibly higher abundance of northern flying squirrels (Meyer et al. 2005, 2007). In northern Sierra Nevada mixed-conifer forests, Innes et al. (2007) found higher densities of dusky-footed woodrats (*Neotoma fuscipes*), another preferred owl prey species, in areas with large black oaks (*Quercus kelloggii* Newberry). In general, these studies suggest that foraging habitat is (1) more open (less vegetation biomass) than nesting habitat, (2) often located close to nest sites, (3) associated with large trees and snags, and (4) infused with other vegetation types (e.g., riparian forests, black oak-dominated patches).

**Spotted Owls and Fire**

The late-successional, and often dense, forests favored by spotted owls for nesting and roosting are at risk to stand-replacing fires because of heavy fuel loading (Agee et al. 2000). Accumulated dead biomass and down woody debris can carry fire horizontally through the forest and vertically into the upper canopy (Tappeiner and McDonald 1996, Weatherspoon and Skinner 1995). Such high fuel loading and vertically continuous ladder fuels put structurally complex, mature forests at greater risk of stand-replacing fire (Agee, 1993, North and Hurteau 2011, Weatherspoon et al. 1992). However, forest landscapes exposed to repeated burning are often
buffered from the effects of future wildfires and characterized by a mosaic of forest patches with high structural heterogeneity at multiple spatial scales (Collins et al. 2009, Stephens et al. 2008). This heterogeneity can improve spotted owl persistence by protecting late-successional patches from stand-replacing fire and potentially enhancing the abundance or diversity of prey species within an individual territory (resulting from greater habitat diversity) (fig. 5-1).

High fuel loading and ladder fuels can reduce foraging or nesting habitat quality for California spotted owls in Sierra Nevada forests. In a fire-suppressed forest, Blakesley et al. (2005) found that increasing the proportion of smaller trees (<23 in d.b.h.) (<60 cm) around the nest, even with high overstory canopy cover (>70 percent), can negatively influence owl occupancy. Decades of fire suppression created a dense understory of regenerating white fir (*Abies concolor* (Gordon & Glend.) Lindley), and these thickets of young trees could interfere with owl foraging in high-use areas. Roberts et al. (2011) found that scattered small trees did not negatively affect owl occupancy in forests where managers allowed low- to moderate-severity fire to periodically clear out these thickets and leave behind large, live trees while retaining high overstory canopy closure. Scattered pockets of small- and medium-diameter trees (4 to 20 in d.b.h.) (10 to 50 cm) can contribute to a multilayered canopy that may allow for efficient thermoregulation for spotted owls, which are not well adapted to heat exposure (Barrows 1981, Weathers et al. 2001). In contrast, Clark (2007) showed that northern spotted owl occupancy and annual survival rates declined (Clark et al. 2011), and annual home range and local extinction increased immediately following (1 to 4 years) wildfire. Clark (2007) also found that annual home range size increased with increasing amount of hard edge (e.g., logging or fire boundaries) within the home range, suggesting lower

Figure 5-1—A California spotted owl nest in a mixed-conifer forest that burned in a prescribed fire with mixed fire severity in 1997 in Yosemite National Park. Note the nest (shown by arrow) is in an area that burned at low severity and has high canopy closure. The nest is adjacent (<50 ft away) (<15.2 m) to an area (left one third of the photo) with lower closure that experienced moderate fire severity.

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quality habitat within fragmented sites. Postfire salvage logging and large areas of early-seral forests in his study area, however, may have confounded his observed occupancy rates.

Fire effects on foraging habitat are not well understood. Clark (2007) observed 23 northern spotted owls using all types of fire severity in southern Oregon. However, within burned areas, owls strongly selected low-severity or unburned areas with minimal overstory canopy mortality. In this burned landscape, owl high-use areas were characterized by lower fire severity and greater structural diversity. Clark (2007) also found that postfire salvage logging reduced owl habitat quality. In contrast, Bond et al. (2009) followed seven owls (three pairs and an individual) using a 4-year-old burned forest in southern Sierra Nevada and found higher than expected owl foraging in high-severity burn areas. The study, however, is limited by its small sample size, brief period of study (12 weeks), and nonrandom owl selection. Additionally, studies of deer mice (*Peromyscus maniculatus*) and other spotted owl prey in Yosemite National Park indicate that deer mouse abundance was negatively associated with increasing fire severity (Roberts et al. 2011). Collectively, these studies suggest the presence of large trees and high overstory canopy closure are the most important pre- and postfire conditions associated with spotted owl occupancy.

References


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